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Anders Ekbom, Gardner M. Brown and Thomas Sterner

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Institutionen för nationalekonomi med statistik Handelshögskolan vid Göteborgs universitet Vasagatan 1, Box 640, SE 405 30 Göteborg 031 786 0000, 031 786 1326 (fax) www.handels.gu.se info@handels.gu.se



Muddy Waters:

Soil Erosion and Downstream Externalities¹

Anders Ekbom, Gardner M. Brown, Thomas Sterner²

Abstract

Soil erosion and fertilizer run-off cause serious flow externalities in downstream environments through-out the world. Social costs include e.g. loss of health, life and production due to pollution and eutrophication of freshwater resources, reduced life of hydro-power plants, increased turbidity, and degradation of coral reefs and marine resources.

The key optimal control models on soil capital management omit downstream externalities and assume that the individual farmer and society share the same objective function. In the presence of externalities, there is a discrepancy. In this paper the social planner aims at maximizing the profits from agriculture subject to a soil dynamics-constraint and external damage costs caused by downstream contamination from soil and fertilizer leakage. These effects are not considered by the farmer.

Comparative statics analysis shows that factors which promote a low discount rate (tenure security, access to credits, crop insurance etc.) will reduce soil erosion and nutrient leakage and promote accumulation of soil capital. Socially optimal subsidies for soil conservation not only will build-up soil capital and increase on-site crop production, but will also reduce nutrient leakage and soil loss. A charge on fertilizer would reduce fertilizer use and thus reduce the water pollution caused by leakage of inorganic nutrients.

Based on our model results, combined with an extended discussion on policy instruments, we conclude that the government should try to provide incentives, not necessarily to stop soil loss per se (since the farmers will look after their own capital) but to avoid contamination of downstream environments, where the resource users have few opportunities to negotiate with the upstream farmers, who may even be unaware of the problems they cause.

Keywords: optimal control theory, micro analysis of farm firms, resource management, soil erosion

JEL classification: C61, Q12, Q20

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² Department of Economics, Göteborg Univ., Box 640, 40530 Göteborg, Sweden anders.ekbom@economics.gu.se

1. Introduction

This article investigates two related questions: (1) what are the likely effects on farmers' land use and downstream environmental quality of government reforms of certain input and output prices and the interest rate, and (2) what policy instruments can successfully address downstream externalities of soil loss and agricultural run-off? Specifically, what are the pros and cons of using policy instruments like information, regulation, taxation of polluting inputs and compensation of farmers for soil conservation via e.g. payments of environmental services (Pagiola *et al.*, 2005; Pagiola, 2007; Smith, 2006; Wunder, 2005)? Our approach is general but the study is particularly important for small-scale agriculture characterized by erosive soils, non-point source pollution and asymmetric information between up-stream polluters and downstream victims. This setting is common in many developing countries.

Soil erosion and agricultural surface run-off cause serious flow externalities in downstream environments throughout the world.³ Pathogens⁴ are carried into water courses and increase morbidity and mortality among downstream water users (Younes and Bartram, 2001). Leaching of soil nutrients cause toxic algal blooms and eutrophication (Anderson 1995; Matson *et al.* 1997; Ayoub 1999), with negative impacts on lake birdlife, fish populations, freshwater resources, and marine ecosystems including coral reefs (Shumway 1990; Horner *et al.* 1997; Naidu *et al.* 1998; Bryant *et al.* 1998; Bartram and Chorus 1999; Ballot *et al.*, 2004). Surface run-off also increases the extent of scours, gullies and floods, which increase the incidence of water-borne diseases (e.g. malaria) and depreciates infrastructure like roads and bridges in downstream areas (White *et al.*, 2000; Fabricius, 2004).

The economics of soil management has a long history and dates back to Wilcox (1938) and Bunce (1942). Significant contributions in this field include papers by Burt (1981), McConnell (1983), Barbier (1990), Barrett (1991), Clarke (1992), LaFrance (1992), Goetz (1997), Grepperud (1996; 1997a,b; 2000), Smith *et al.* (2000) and Yesuf (2004). Soil is natural capital and needs to be managed as an integral part of the farmer's (or social planner's) objective function to maximize the long run private (or social) net profits from agricultural production. In the analytical formulation of this problem, the researcher can assume, as we do, that a farmer uses resources to enhance soil properties, thereby making it a renewable natural resource.

³ Soil erosion and surface-run off also cause a set of negative *stock* externalities. These include e.g. sedimentation of water reservoirs, hydro-power plants, irrigation and other fresh-water supply structures, river estuaries (build-up of mud banks), and coastal and marine environments, including corals reefs. Although stock externalities can be important we focus in this paper on flow externalities.

⁴ Viruses, bacteria and helminths (e.g. roundworm, whipworm, and hookworm).

All of these studies have a concern for the loss of the natural capital that soil represents to the farmer. However, none of the economic studies cited above have focused on the off-site externalities, yet the associated social costs are significant. Smith (1992), for instance, reports that the mean annual off-site damage cost⁵ to US agriculture due to flow externalities amounts to 4.6 % of the value of that sector's output. In erosive tropical areas the damage could be higher. Moreover, there are places in the tropics where soil erosion/conservation is fairly low on the farmers' agenda since they may have very deep fertile soil but hardly any other assets. Soil erosion can however still be a large problem for people living downstream.

We contribute to the literature by developing a model which incorporates the downstream social consequences of upstream private decisions. We further discuss appropriate policies for managing offsite effects such as regulation, taxation, subsidies or markets for ecosystem services. The article is organized as follows. First, we present a simple generic optimal control model of crop production with flow externalities and soil dynamics. Second, we analyse comparative statics of the model by identifying and discussing effects of changes in some policy variables and discuss potential policy instruments. Finally, we summarize and discuss our findings and draw some policy conclusions.

2. An Optimal Control Model of Soil Management with Downstream Damage

Assume that agricultural production is determined by the following production function:

(1)
$$Q = f(S, L_Q, F)$$

where agricultural output (Q) is a function of soil capital (S), labour supply to agricultural production (L_Q), and chemical fertilizer (F). Output may consist of the value of one or several crops. Although soil is a heterogeneous resource, which consists of several properties, the present model treats soil as a single, one-dimensional variable. While recognizing that soil capital consists of a range of biological, physical and chemical properties⁶, soil depth is critical for adequate root-holding capacity and other soil properties necessary for good plant growth (Thomas 1994). Let (S) represent an overall index of soil capital. It is an abstraction, but serves as a proxy for the soil properties, which make up the total

⁵ External costs pertaining to freshwater and marine recreation, water storage, navigation, flooding, irrigation, commercial fishing, municipal water treatment, and municipal and industrial use.

⁶ For instance, macro nutrients (e.g. nitrogen, phosphorus, potassium), micro-nutrients (e.g. copper), cat-ion exchange capacity, moisture, permeability, structure, clay-sand-silt content and pH-level. See Ekbom (2007) for further discussion of the many dimensions actually involved in S.

capacity of soil to produce output. $f(S, L_Q, F)$ is assumed to be well-behaved⁷. Specifically, in order to identify the effect of changes in policy parameters on the steady state values of the key variables we assume that f() is concave; it is increasing in each of its arguments: $f_S > 0$, $f_{L_Q} > 0$, $f_F > 0$ (the subscripts indicate the partial derivative with respect to the variable) and subject to diminishing marginal returns, $f_{SS} < 0$, $f_{L_QL_Q} < 0$, $f_{FF} < 0$. The Hessian matrix of $f(S, L_Q, F)$ is negative definite: $f_{LL}f_{SS} - f_{LS}^2 > 0$, $f_{SS}f_{FF} - f_{SF}^2 > 0$, $f_{LL}f_{FF} - f_{LF}^2 > 0$ and $f_{LL}f_{SS}f_{FF} + 2f_{LS}f_{SF}f_{LF} - f_{SS}f_{LF}^2 - f_{FF}f_{LS}^2 - f_{LL}f_{SF}^2 < 0$. We also assume that $f_{ij} > 0$; $i, j = S, L_Q, F$; $i \neq j$.

The typical setting for our model is a developing country where small-scale farming is practiced on steep slopes under erosive tropical rains. The cultivation is not mechanized and depends on family labour. We assume technology to be constant. The household's main cash expenditure on farming inputs includes chemical inorganic fertilizers, used to boost crop production and compensate for nutrients losses due to soil loss.

We introduce the following soil dynamics:

(2)
$$\dot{S} = g(L_c) - \psi(L_o) + \sigma$$
,

where change in soil capital, $dS/dt = \dot{S}$ is a function of labour supply to soil conservation (L_c) , to agricultural production (L_Q) plus the natural rate of net soil accretion or erosion, σ . Based on empirical evidence, it is reasonable to assume that $g'(L_C) \ge 0$, $g''(L_C) \le 0$, $\psi'(L_Q) \ge 0$ and $\psi''(L_Q) \ge 0$. Labour used for soil conservation is assumed to build up soil capital, although at a diminishing rate. Labour used for cultivation is assumed to depreciate soil capital. Cultivation practices like plowing and seed-bed preparation typically break the soil's physical structure, accelerate volatilization of nutrients, and increase the soil's susceptibility to erosion (Morgan 1986; Troeh *et. al.* 1991; Thomas 1994). Common to many developing countries, the markets for labour are local but functioning. Hence, we assume separability between L_Q and L_C . An additional assumption is that $\sigma = 0$, which implies that natural soil accretion and natural soil erosion balance out to be zero or

⁷ Focus in this paper is not on stability or uniqueness of equilibria, nor are we interested in special cases such as corner solutions. We assume functions sufficiently well-behaved to give interior solutions.

negligibly small in the relevant time period. The latter assumption is an approximation but may be reasonable given two facts: first, natural soil accretion is a very slow process; second, soil loss on virgin lands is very small⁸.

To operationalize the distinction between the farmer's and the social planner's objective function-and focus on the point that soil erosion and surface run-off cause substantial downstream damage, we introduce the following cost function that captures the relationship between downstream environmental quality and soil dynamics:

(3)
$$E = b[S - \Phi(F)] = b[g(L_c) - \psi(L_o) + \sigma - \Phi(F)]$$

.

in which downstream environmental quality (*E*) is a function of the flow of eroded soil (*bS*) b>0, the net soil accretion, and run-off (or leaching) of chemical fertilizers ($\Phi(F)$). *E* is a placeholder for offsite damages to the quality of downstream environmental resources like rivers, lakes and reservoirs used for drinking-water supplies, marine coastal waters and coral reefs. Following our earlier assumptions, $E_{L_c} > 0$, which implies that enhancing the soil's physical, chemical and structural properties through soil conservation reduces the risk of soil erosion and downstream damages. This is in accordance with research findings by e.g. Troeh *et. al.* 1991. Moreover, a marginal increase in labour supply to agricultural production increases soil erosion, and increases the flow externalities of suspended soil particles in downstream water resources ($E_{L_0} < 0$), and increased use of chemical fertilizers contributes negatively to the quality of downstream water resources due to surface run-off ($E_{E} < 0$).

Given a certain technology, the social planner's objective function is to maximize the discounted net social profit (π) from agricultural production over an infinite time horizon⁹:

(4)
$$\pi = \int_{t=0}^{\infty} \left[pQ - w(L_{c} + L_{Q}) - vF + b(\dot{S} - \Phi(F)) \right] e^{-rt} dt .$$

of the social function analysed for the value b=0.

⁸ Mature forest-, bush- or grass-lands typically offer very dense ground cover and cause minimal soil loss. It is cultivation that breaks up the soil and triggers the accelerated soil erosion process (for a comparison between soil loss on natural lands and bare (cultivated) plots see e.g. Thomas, 1994, Table 5.6, p. 144).

⁹ The profit function of the private farmer (π_p) takes the following form:

 $[\]pi_{P} = \int_{t=0}^{\infty} \left[pQ - w(L_{C} + L_{Q}) - vF \right] e^{-rt} dt$. Both profit function and its solution can be seen as a special case

(p), (v), (w) and (r) are given parameters representing the price of output, fertilizer, labour and the discount rate, respectively.

Using Pontryagin's Maximum Principle (Pontryagin *et. al.* 1964), maximizing equation 4 subject to equations 1-3 is done by maximising the following current value Hamiltonian (H):

(5)
$$H = pf(S, L_{\varrho}, F) - w(L_{\varrho} + L_{c}) - vF + \lambda (g(L_{c}) - \psi(L_{\varrho}) + \sigma) + b (g(L_{c}) - \psi(L_{\varrho}) + \sigma - \Phi(F)),$$

where λ is the co-state variable.

Assuming an interior solution, the first order necessary conditions for equation 5 are:

(6)
$$\frac{\partial H}{\partial F} = 0 \implies pf_F = v + b\Phi'(F),$$

(7)
$$\dot{\lambda} - r\lambda = -\frac{\partial H}{\partial S} = -pf_S$$

...

(8)
$$\frac{\partial H}{\partial L_{Q}} = 0 \implies pf_{L_{Q}} = w + (b + \lambda)\psi'(L_{Q})$$
, and

(9)
$$\frac{\partial H}{\partial L_c} = 0 \implies (b+\lambda)g'(L_c) = w$$
.

The necessary conditions have familiar interpretations. Equation 6 requires factor market equilibrium; the value of the marginal product of fertilizer (pf_F) should equal its private marginal cost (v) plus the marginal social downstream cost of fertilizer use $(b\Phi'(F))$. Rearranging equation 7 into the following expression: $r = \dot{\lambda}/\lambda + pf_S/\lambda$ yields the standard arbitrage equation in capital theory, where the competitive rate of return earned for holding any other asset of equivalent risk (r) should at all times equal the return on soil capital due to price appreciation or depreciation $(\dot{\lambda}/\lambda)$ plus the real yield from soil capital in production (pf_S/λ) .

Equations 8 and 9 introduce some new information pertaining to downstream flow externalities compared to earlier studies on optimal soil use. According to equation 8, the value of the marginal

product (VMP) of labour in agricultural production (pf_{L_Q}) should in equilibrium equal the market wage rate (w) plus two marginal contributions: downstream flow damages from cultivation labour $(b\psi'(L_Q))$ and the shadow value of soil depletion $(\lambda\psi'(L_Q))$. Equation 9 implies that the marginal social downstream benefit of soil conservation $(bg'(L_C))$ plus the marginal effect on *in situ* soil capital of conservation $(\lambda g'(L_C))$ should in equilibrium equal the market wage rate (w).

In steady state equilibrium, when neither stocks nor prices change, $\dot{S} = \dot{\lambda} = 0$. Then from equation 2,

(10)
$$g(L_c) + \sigma = \psi(L_o),$$

which implies that soil conservation and the labour devoted to it, adjusted for natural changes(σ), should be sufficient to offset loss of soil capital from cultivation. Moreover, in steady state the sign of $\frac{dL_c}{dx}$ equals the sign of $\frac{dL_Q}{dx}$ (where x = r, w, v, p) since by total-differentiating equation (10) above we get $dL_c = \frac{\Psi'}{g'} dL_Q$.

Further, in steady state equilibrium, according to equation (7), (11) $\lambda = \frac{pf_s}{r}$,

which says that the rental rate of soil capital (λ) should equal the capitalized value of the productive future use of this soil (pf_s / r).

3. Comparative Statics – Results and Interpretation

Using comparative statics we derive how marginal changes in policy parameters affect some key variables relevant to the farmer's production as well as the flow externalities. The policy parameters considered are the interest rate (r), wage rate (w), fertilizer price (v), and crop price (p). The derivations of the comparative statics results are contained in Appendix 1 and are summarized in Table $1.^{10}$

¹⁰ The results apply for a general production function. By imposing restrictions, further results can be obtained. For example, a Cobb-Douglas production function implies that dF/dp>0.

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	Effect on				
Change in	Soil (dS)	Labour (dL _C ; dL _Q)	Fertilizer (dF)		
Interest rate (dr)					
	< 0	< 0	< 0		
Wage rate (dw)					
	?	< 0	?		
Fertilizer price (dv)					
	?	< 0	< 0		
Crop price (dp)					
	?	> 0	?		

? = Sign undetermined

The most transparent and unambiguous comparative statics results arise from a change in the interest rate. Similar to the findings in e.g. McConnell (1983), Barrett (1991) and LaFrance (1992), when there is a permanent and unanticipated increase in the interest rate, soil capital is reduced because increasing returns on rival capital require disinvestment in soil capital in order to increase its marginal productivity. A difference in the results obtained in this study compared to earlier similar studies, indicate that the inclusion of off-site impacts in the objective function reinforces this effect. In other words, a reduction in the interest rate will result in indirect additional benefits in terms of reduced downstream externalities. Note that these effects are tied to soil stock changes that only occur along the transitions from one steady state to another.

Naturally, factor demand decreases when its own price increases. Whether factor demand increases or decreases when another factor price increases depends on the strength of the substitution effect increasing demand compared to the output effect decreasing factor demand. The net result depends on the production technology. In the presence of a strictly concave Cobb-Douglas production technology, the quantity of factor *i* decreases when the price of factor *j* increases because the output effect dominates the substitution effect. In our case the comparative static results are made more complicated by the feedback phenomenon induced by the soil dynamics equation (2). The sign for dS/dw, dS/dv and dF/dw is ambiguous for the following reason. When the wage rate changes, say increases, a decrease in L_Q decreases soil loss and increases *S*. It also decreases conservation labour and *S* along with it. Without putting further structure on the technology, the sign of dS/dw is therefore

indeterminate. More quantitatively, as can be seen in Appendix 1 (eq. 20), the sign of $\frac{dS}{dw}$ is

ambiguous since we cannot determine *a priori* whether
$$\left(\frac{g''}{g'}\Psi'+\Psi''\right)>0$$
 or ≤ 0 . $\frac{dS}{dw}<0$ if

 $\frac{g''}{g'}\Psi' + \Psi'' \ge 0 \quad \text{or} \quad \left|\frac{g''}{g'}\right| \le \left|\frac{\Psi''}{\Psi'}\right|, \text{ i.e. if soil conservation labour exhibits less curvature than the}$

negative impact on soil capital of cultivation labour.

There is a similar effect for dS/dv. When the price of fertilizer changes, say increases, the substitution and output effect play out in some fashion with respect to S and L_Q . However, the change in L_Q causes S to move in the opposite direction via the soil dynamics equation and further contributes to the indeterminacy we observe.

The feedback forces, due to the soil dynamics equation, contribute further to the ambiguous sign of dF/dw. Should w increase, there is the negative output effect together with the positive substitution effect for S and L_Q . These changes transmitted to the soil dynamics equation individually influence S in an indeterminate fashion, which then affects the endogenously determined shadow price of S in an indeterminate manner. How F ultimately equilibrates is affected by the new price ratio, λ/w . Similarly, the sign of the effect of an increase in crop price on fertilizer use is undetermined, since we

cannot sign $(g''/g')\Psi' + \Psi''$. However, a wage increase negatively affects fertilizer use $\left(\frac{dF}{dw} < 0\right)$

if $\frac{g''}{g'}\Psi' + \Psi'' \ge 0$ or $\left|\frac{g''}{g'}\right| \le \left|\frac{\Psi''}{\Psi'}\right|$ (for details, see eq. 17 in the Appendix). As before, the wage effect

is positive if soil conservation labour is less elastic than cultivation labour.

dS/dp is ambiguous because when productive labour increases because *p* increases, this decreases *S* in the soil dynamics equation, that offsets to an unknown amount the positive effect of a positive product price change on *S*. Although his model assumptions are slightly different¹¹, Barrett (1991) obtains a similar result. He finds that the sign of the effect on soil conservation and soil depth of an increase output price is indeterminate, unless one makes specific assumptions about the technical relationships and dependence between soil, soil loss attributable to cultivation, soil conservation and non-soil inputs (viz. fertilizers).

¹¹ For instance, Barrett uses a Cobb-Douglas production function, and assumes that farmers choose the amount of soil loss directly in their production; the cost of labour is not included.

A relevant question which follows from the comparative statics results is why there is a negative effect of fertilizer price (v) on labour use? Arguably, the result is created by two effects. First, we have two opposing forces: as v increases there is *substitution* out of fertilizer into the other factors, so labour use goes up and soil capital (S) should go up too. Familiarly, the *output effect* caused by the fact that fertilizer is now more expensive induces labour to decrease and S should go down too. Second, S too changes through the feedback in the soil dynamics equation through changing labour use (both cultivation and conservation). Apparently, the output effect, combined with the soil dynamics feedback effect, dominates the substitution effect.

Policy Instruments to Mitigate Downstream Effects

Most analyses of soil loss have a limited focus on policy instruments, which address on-farm concerns. Given our model and the comparative statics results, we discuss policy instruments below in an environment where there are off-site externalities. The key question facing the social planner is thus: what (mix of) policy instruments enables the government to maximize the discounted social profit from agricultural production subject to downstream externalities caused by soil erosion and fertilizer run-off. The policy maker may choose between a large set of policy instruments such as (i) direct regulation, (ii) information, (iii) property rights, (iv) charges and (v) subsidies¹². In the choice of relevant policy instruments it is important to also consider issues regarding rights, fairness (distributional and equity concerns), efficiency and administrative feasibility (Sterner, 2003). Although the specific (historical, social or political) context may prevent real implementation of some policy instrument(s) presented below, it is nevertheless possible and useful to discuss the experiences and the pros and cons of these instruments in a developing country perspective.

(*i*) *Direct regulation:* Theoretically, direct regulation would imply that farmers were obliged to supply cultivation labour, fertilizer and soil conservation labour corresponding to their respective socially optimal levels (given by eq. 6, 8 and 9). Although the privately and socially optimal levels of soil conservation differ, governments have frequently used direct regulation (in terms of cultivation bans, certain soil conservation requirements etc.) as a policy instrument to address soil erosion and run-off (Hudson 1981; Morgan 1986). Kenya is no exception in this respect. Soil conservation was made compulsory on cultivated land in 1937. Until Independence in 1963, implementation of soil conservation among the native African farmers relied on government orders, regulation, coercion and

¹² Due to lack of practical experiences, the complexities and the substantial institutional requirements associated with use of other policy instruments, like tradable permits (for reference, see Sterner (2003)), they are not considered in this paper.

penalties. Mandatory engineering solutions, such as construction of labour-intensive bench terraces, cut-off drains, stone gabions and retention ditches, were prescribed (Kimaru 1998).

Although the choice and implementation of policies have changed considerably since Kenya's independence, regulation is still an important element of the country's soil conservation efforts. Farmers are required by law to conserve their soil. Based on specific soil conservation requirements for different types of land, the local soil conservation officer keeps records of what soil conservation measures individual farmers have to establish. Failure to establish these measures subjects them to an elaborated set of graduated sanctions. Other examples of regulatory command and control measures pertaining to soil use in Kenya include bans on cultivating soils above certain hill slopes (>60%) or along river-banks, or vertical ploughing (perpendicular to the contour). Due to population pressure, lack of knowledge, insufficient enforcement and other reasons, these bans are frequently violated.

However, the regulatory approach to soil conservation has largely been unsuccessful. The underlying cause can be found in the farmer's incentive structure. Our model shows that a privately rational farmer would only conserve soil up to the point where the marginal benefit on in situ soil capital of conservation $(\lambda g'(L_c))$ equals the market wage rate (w). In the normal case, the marginal social downstream benefit of soil conservation $(bg'(L_c))$ will not be internalized in the farmer's economic decision. In other words, a poor farmer who cultivates deep fertile soils on steep slopes and is constrained in labour and cash, has for rational reasons small incentives to prevent all soil loss and fertilizer run-off from his/her land. Conserving all soil implies that the farmer will bear the full social cost of soil conservation and preventing downstream damages, whereas only a share of the benefits accrue privately. Since the marginal social downstream benefit of soil conservation $(bg'(L_c))$ is essentially public, a rational resource-constrained farmer will not (or cannot be expected to) pick up the cost of attaining it. Similarly, poor farmers cannot be expected to prevent the public downstream flow damages $(b\psi'(L_0))$. Thus farmers continue to produce public bads in terms of degradation of downstream water resources, siltation, sedimentation and pollution. In contrast to the individual farmer's financial reasons, the social planner has a strong economic reason to encourage full soil conservation, discourage soil erosion and prevent downstream damages.

(*ii*) *Information:* Increasing the knowledge among farmers has frequently been used by governments to promote sustainable agriculture. In Kenya, this has been pursued by disseminating the benefits of soil conservation and costs of soil loss, and provision of practical extension advice to small-scale farmers on how to conserve soil and attain sustainable land husbandry. These activities have largely

replaced earlier land use-policies based on coercive regulation. Since 1974, farmers have been offered specific SWC field training, study visits to research stations, on-farm advice by soil conservation extension officers and educational material in a National Soil and Water Conservation Programme (NSWCP). Farmers have been organised into Catchment Planning Teams with the purpose of conserving soil in a coherent manner in designated geographical areas (Admassie 1992; OPTO, 2006).

In general, information can be a cost-effective policy instrument for environmental management (Sterner, 2003). Kenya's government's use of information to increase soil conservation implementation has been rated rather successful (OPTO, 2006; Kimaru 1998; Lundgren 1993)¹³ and Kenya's farmers have voluntarily increased their soil conservation efforts, quantitatively as well qualitatively. However, downstream damages due to soil loss and fertilizer run-off remain a large problem. This indicates that traditional information on soil conservation technologies is a useful but insufficient policy instrument to *fully* prevent soil erosion and downstream damages.

The reason to this lies in the individual farmer's objective function. The farmer's objective is to maximize private discounted profits (π_p) without considering external effects. It is true that L_c and L_q embody skills obtained inter alia from the government's NSWCP but this knowledge mainly assists farmers to fulfil their private objectives. Hence, increased information cannot be expected to produce *socially* optimal outcomes. In other words, π_p excludes off-site damages and increasing the amount of information to farmers does not alter the fundamental economic incentives driving their behaviour. Another complication regarding information is the fact that identifying and disseminating the specific downstream effects caused by individual farmer's agricultural production is also very difficult in cases characterized by non-point source pollution and geographically remote externalities.

A complementary policy instrument would be to increase farmers' knowledge on public investments and development plans. Our comparative statics result regarding the interest rate (dS/dr < 0) suggests that increasing farmers' knowledge on public measures, which reduce farmers' discount rate (e.g. input-/output market development plans, road investments, land tenure reforms etc.), has positive effects on soil capital formation and indirectly prevents downstream damages. Moreover, information on downstream effects from agriculture is highly relevant for the social planner in fulfilling its

¹³ The positive effects of extension advise have been contested by Evenson and Mwabu (2001) and Gautam and Anderson (1999), who found limited evidence of significant positive effects on farmers' agricultural productivity of Kenya's Training and Visit system for agricultural extension services.

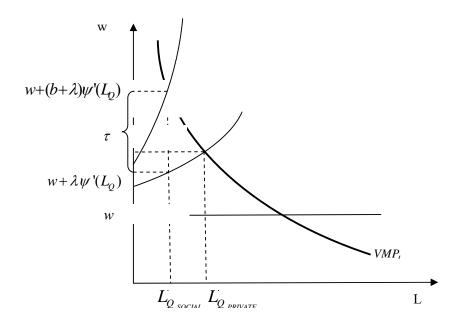
objective function, and in the design of economic policy instruments (such as charges, fees, subsidies), which can be used to curb the externalities.¹⁴

(iii) Charges or fees: In principle, the external costs imposed on downstream victims should be internalized into the farmers' production costs. The Pigouvian approach would be to put a charge or a fee on the degrading inputs (or practices). Illustrated in figure 1 below, a rational farmer (with secure rights) would use cultivation labour such that VMP of L_Q equals the market wage for labour plus the marginal effect on soil capital $(\lambda \psi'(L_Q))$. This corresponds to $L_Q^*_{p_{RNATE}}$. However, since cultivation labour depreciates soil and cause downstream flow externalities (represented by $b\psi'(L_Q)$), it is socially optimal to reduce the use of cultivation labour to $L_Q^*_{p_{SOCML}}$. Reducing erosive cultivation labour could in principle be achieved by coercive measures e.g. restrictions on how much labour one can use (on a given plot of land) for agricultural production with, however, the attendant problems of monitoring and enforcement.

Introducing textbook economic incentives, one can instead introduce a charge, τ , corresponding to $b\psi'(L_Q)$ in equation 8. In practice, this is however also hard to enforce but some more realistic policies to manage downstream externalities are discussed later.

¹⁴ In special cases where *payments for ecological services* (see sub-section (v) below) may be obtained, information of downstream costs of soil loss or social benefits of soil conservation, may be strategically important knowledge to individual farmers as well in order for them to take advantage of this financial benefit.

Figure 1. Agricultural labour demand – the effect of a pollution charge



Regarding fertilizer, from (6) we know that a privately rational farmer would use fertilizer in such an amount that VMP_F equals the fertilizer price $(pf_F(S, L_Q, F) = v)$. However, since fertilizer use also produces a negative externality $(b\phi'(F))$, the government ought to introduce a charge or a fee which internalizes this social cost. Would this be a viable policy instrument to achieve the social planner's objective function? As shown by the comparative statics results (in Table 1), raising the farm-gate price of fertilizer, through a charge or a fee, reduces fertilizer use $\left(\frac{dF}{dv} < 0\right)$ and thus the nutrient run-off into water systems.

However, a charge on downstream pollution is problematic for several reasons. Firstly it is politically very sensitive. Farmers may be rich and powerful or – as in many tropical countries so poor that they can hardly support additional taxation. In principle it might be possible to construct a package in which increased fertilizer taxes are counteracted by lowering other taxes for instance on their output. Introducing a tax or a charge on erosive cultivation would however also be infeasible for monitoring and enforcement reasons. Soil erosion is typically a non-point source pollution problem, which originates in vast watersheds and is caused by thousands or even millions of small-scale farmers' agricultural production. A pure downstream pollution tax would be infeasible since there is insufficient monitoring ability. Joint schemes to make farmers collaborate in reducing pollution are

possible but much more complex. One component of the pollution – that which comes from commercial fertilizers - could of course be taxed. Irrespective of whether a fertilizer charge is targeted at farmers by an *ad valorem* tax or directly at the producers, a fertilizer charge increases the farmer's production costs, reduces their profits, and will therefore be severely resisted. It may also be thought of as running counter to policies designed to improve food security and self sufficiency.

It may be argued that a fertilizer tax can be used to subsidise conservation labour (L_c) ? Combining these two policy instruments is however difficult basically for reasons of efficiency. Recall that our model says: i) to tax fertilizer use a very specific amount to achieve efficiency, ii) there is a very specific amount of subsidy for L_c to achieve efficiency. To ensure efficiency, one has to keep these two policies separate. In practice that may be difficult. If you explicitly tie one policy to the other, then farmers have an incentive to distort their behaviour. The farmer might strategically use more fertilizer (which increases private yield but causes downstream damages) in order to increase the "subsidy fund" for his/her conservation labour. So formally one has to keep efficiency decisions separate from financing decisions.

(*iv*) *Property rights:* Enhancing property rights is a policy instrument, which implicitly reduces farmers' discount rate. Land ownership security affects both investment incentives and the availability of resources to finance investments (Feder and Feeny 1991). Farmers holding title deeds to their land may use it as collateral for credits, which enable land investments such as terracing or tree plantation. In a case study of northern Ethiopia, land tenure security was positively associated with soil conservation investments (Alemu 1999). Feder and Onchan (1987) find that land-improving investments are positively affected by ownership security.

Recalling the comparative statics results, reduced interest rates builds up soil capital (dS/dr < 0), increases labour supply to soil conservation ($dS/L_C > 0$) and thus reduces downstream externalities. In Kenya's Central Highlands land tenure security is relatively high compared to e.g. its neighbouring countries. A majority of the smallholders hold title deeds to surveyed, registered and adjudicated land, which can be sold or bought in an open market. The title deeds prevent arbitrary evictions and facilitate bank loans. However, as land fragmentation accelerates due to population growth and sub-division of farms, the government has an important role to play. Traditionally, land is owned by men and inherited by sons. Women who head households, divorced women or widows enjoy no rights to hold land or obtain a title deed to their specific plot. Consequently, they have little incentives to invest in land they cultivate. This introduces distortions in the land market and reduces the tenure security. An important policy measure is thus to adjust the current institutions governing land ownership with

respect to the existing distortions, and e.g. facilitate registration of sub-divided land and strengthen womens' rights to own, buy and sell land, and use land as collateral for credits.

Strengthening on-farm tenure security is necessary but as our model shows insufficient to fully prevent downstream externalities. A complementary measure would be to strengthen the human right of downstream inhabitants to clean water. These users' right to clean water needs to be acknowledged, formally defined, clarified and enforced. This implies a responsibility on the government to increase the provision of clean water, through e.g. intensifying the support to soil conservation, decontamination of existing water sources, redistribution among existing users/sectors, and/or increasing the supply from other freshwater sources.

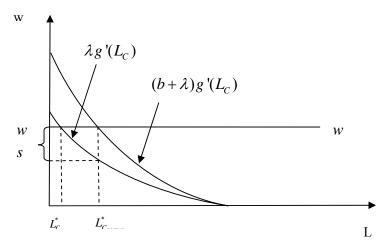
Regarding <u>equity and rights</u> the critical question is who is entitled to what right? Are downstream water users entitled to clean water, or do the upstream farmers hold the right to pollute? It seems natural to argue that all downstream victims should be compensated (by the polluters) for the damage inflicted on them. However, the problems pertaining to soil erosion, sedimentation and nutrient leakage are typically characterized by asymmetric information, and direct compensation between all polluters and victims implies very high transactions costs. Moreover, in Kenya's highlands and in many other similar situations in developing countries, up-land farming started long before downstream hydro-power production, irrigation and coastal tourism were initiated. Although soil conservation was mandatory in Kenya (and other East African countries) under the British colonial rule, it was never fully enforced. It has become more and more an accepted fact that soil loss naturally occurs as an unintended negative side-effect of resource-constrained small-scale farming on erodible soils in tropical hilly environments. The farmers can thus claim a historical prescriptive right to pollute. It may thus be argued that the more recent downstream economic activities (hydro-power, irrigation etc.) had an obligation - prior to their investments - to properly internalize the cost of environmental inputs (including polluted water) in their production and ensure adequate protection against it.

Regarding poor people, who reside in the downstream areas and depend on the water resources for their livelihood, the equity and rights issues lead to another conclusion. This group is financially and politically much weaker than the hydro-power companies, tourism operators etc. Typically, they settled in the low-land area before the high-land farmers settled in theirs (Ochieng and Maxon 1992). As farming has become more intensive and expanded into virgin mountain forests, sedimentation of the water resources on which they depend has increased. Hence, unanticipated at the time of settlement they have become victims of increasing water pollution. As opposed to the hydro-power companies

and other commercial operators, they lack capital for pollution protection and prevention. It may thus be argued that they are entitled to some compensation.

(v) Subsidies and Payments for Environmental Services: Subsidies have the advantage of introducing a positive incentive to encourage a desirable action. As illustrated in figure 2 below, a competitive farmer would build up soil by using soil conservation labour in an amount such that the private marginal value of conservation labour $(\lambda g'(L_c))$ equals the market labour wage rate (w). This corresponds to $L_{C_{PRVATE}}^*$, which however is too little to prevent downstream flow externalities. A farmer who behaves altruistically and conserves more soil than the privately optimal amount produces environmental public goods to society $(b g'(L_c))$. For society to encourage soil conservation up to the socially optimal level $(L_{C_{SOCIML}}^*)$, the farmer would need some form of compensation or a financial transfer (s), which corresponds to this level.





Historically, subsidies to soil conservation have primarily been provided to prevent private yield losses¹⁵. Given the negative externalities inflicted on downstream populations, the government may create new property rights and decide that the downstream population has the right to clean water, the coastal population has the right to coral reefs, etc. From these rights *Payments for Environmental Services* (PES) ensue.

¹⁵ For example, Kenya's government has provided subsidies in kind (e.g. tree seedlings, tools, implements) and cash payments to encourage farmers to conserve soil in order to maintain crop yields and sustain food self-sufficiency.

PES have emerged as a new innovative policy instrument to encourage watershed management and reduce downstream externalities (Smith, 2006; Pagiola and Platais, 2002; Gutman, 2003, Pagiola *et al.*, 2005; Pagiola, 2007, Kerr, 2002; Landell-Mills and Porras, 2002; Wunder, 2005). PES have also been found to be an effective instrument for upstream-downstream problem resolution (Kosoy *et al.*, 2006). Essentially PES is a subsidy, but ensues from established property rights and presupposes a broader (social) scope to soil erosion and soil conservation. In our case, provision of PES implies that farmers who conserve soil are compensated for public environmental services¹⁶ they provide to society.

Although a soil conservation subsidy in terms of PES does not cause the same win-lose effect as fertilizer charges¹⁷, it has both pros and cons: in our case, PES might work if it functions as a real incentive for farmers to conserve soil beyond the privately optimal level $(L^*_{C_{PRVATE}})$ up to the socially optimal level $(L^*_{C_{SOCALL}})$. The social costs are mainly associated with the social revenues necessary to cover the payments. Subsidies increase the government's public expenditures and therefore have to be used with care. This is particularly relevant in developing countries with a very constrained budget. For PES to function efficiently, successful implementation requires monitoring and enforcement due to the inherent risk of free riders (some farmers are paid for services not provided). PES may work in situations where the incentives are compatible for both service users (downstream victims) and service providers (upland farmers), where tenure security is high, transaction costs are low, and where the benefits of the environmental services equal or exceed the costs to the service providers (Landell-Mills and Porras, 2002; Pagiola *et al.*, 2002, 2005). Other critical issues in implementing PES include (i) the characterization of the ecological services, (ii) the establishment of sustainable financing mechanisms, (iii) the design and implementation of effective payment systems, and (iv) the establishment of adequate institutional frameworks (Campos *et al.*, 2005; Sierra and Russman, 2005).

4. Summary and Conclusions

Agricultural production pursued by small-scale farmers on hillsides of developing countries commonly causes downstream damages due to soil erosion and nutrient run-off, which reduce society's total welfare. This problem is addressed in an optimal control model, in which a social

¹⁶ Environmental services such as protecting freshwater quality, controlling hydrological flows, reduced suspension and sedimentation of water systems, prevention of floods and landslides, biodiversity conservation and carbon sequestration.

¹⁷ Pollution reduction is attained at the expense of reduced crop production.

planner maximizes the social profits from farmers' agricultural production subject to external damage costs and a soil dynamics-constraint. These downstream effects, omitted in other formal models, are substantial and presuppose that the individual farmer and the social planner share the same objective function. In our case with externalities, this is not true.

In the world of a strictly concave production technology and all factors are substitutes, many of the comparative statics results are routine. Levels of the factors except soil, vary directly with product price and indirectly with own price. Factor demand varies inversely with an increase in the discount rate. Therefore, factors which promote a low discount rate (tenure security, access to credits, crop insurance schemes) are likely to reduce soil erosion, build up soil capital and prevent water pollution from fertilizer run-off.

We expect the output effect of a factor price change to dominate the substitution effect but the results are ambiguous for changes in soil quality induced by changes in the wage rate or fertilizer price and for the impact of a wage change on fertilizer use. The ambiguity arises because of feedbacks stemming from the equation governing soil dynamics. For example, a wage increase should decrease soil capital, but a decrease in productive labour also reduces the intensity of cultivation and increases soil quality.

Further, the analysis shows that an increase in fertilizer price is negatively associated with fertilizer use, and conservation and cultivation labour, respectively. This suggests that a charge on fertilizer yields mixed effects with respect to downstream externalities: a fertilizer charge (i) reduces fertilizer use and thus reduces water pollution from nutrient run-off, and (ii) reduces both soil conservation and labour supply to cultivation. Without further model assumptions, the net impact of (ii) on on-site soil capital or downstream environmental quality cannot be determined *a priori*.

The results also show that an increase in crop price is positively associated with labour supply to soil conservation and cultivation. From the perspective of downstream effects, this result may be interpreted in at least two ways. First, increased soil conservation will build up soil capital and reduce loss of nutrients. Second, increased crop prices will boost the supply of cultivation labour which will accelerate soil loss. Due to these opposite effects on soil capital and downstream damage, it is difficult *a priori* to establish the impact of changed crop prices. If one can establish empirically, that the positive effects dominate, the government ought to increase (implicitly) the farm-gate selling prices by investing in feeder-roads and other factors which reduce farmers' transport and marketing costs.

Due to the ambiguous results of changing the crop and fertilizer prices, we argue that payments for environmental services, targeted at up-stream soil conservation, should be encouraged. Provided that these payments can be financed and enforced, PES would reward socially optimal behaviour by providing incentives to build up private soil capital (which increases output and the value of the land) *and* produce environmental benefits to downstream (water) resource users.

Based on our findings we conclude that the government may play a crucial role in defining appropriate policies and implementing reforms which encourage farmers to maximize *society's* profits from agricultural production, build up soil capital, prevent soil erosion, and counteract downstream externalities from soil loss and nutrient leakage. Government reforms, which aim at boosting crop production and make use of policy variables like agricultural input prices, crop prices and the interest rate, need also to consider their external downstream effects.

Appendix 1. Comparative Statics Analysis

Use equation (11) in the text to substitute for λ in (8) and (9):

(8')
$$pf_L(S, L_Q, F) - w - \left[b + \frac{p}{r}f_S(S, L_Q, F)\right]\Psi'(L_Q) = 0$$

(9') $\left[b + \frac{p}{r}f_S(S, L_Q, F)\right]g'(L_C) - w = 0$

Total differentiation of equations (6) (8') and (9'), and total-differentiating equation (10) in the text $(dL_c = \frac{1}{g'} dL_Q)$ used to substitute for dL_c) yields the following system:

$$\begin{aligned} I = \begin{bmatrix} y_{FF} \\ y_{FF} \\ z_{F} \\ z_$$

Comparative Statics of the equation system represented by (15) using Cramer's rule is given by (17-28) below:

$$(17) \quad \frac{dS}{dr} = \frac{\begin{bmatrix} 0 & pf_{FL} & [pf_{FF} - \Phi''] \\ -\frac{p}{r^2} f_S \Psi' & \left[pf_{LL} - \left(b + \frac{p}{r} f_S \right) \Psi'' - \frac{p}{r} f_{SL} \Psi' \right] & \left[pf_{LF} - \frac{p}{r} f_{SF} \Psi' \right] \\ \frac{p}{r^2} f_S g' & \left[\left(b + \frac{p}{r} f_S \right) \frac{g''}{g'} \Psi' + \frac{p}{r} f_{SL} g' \right] & \frac{p}{r} f_{SF} g' \\ |J| < 0$$

since the sign of the numerator is negative:

$$\frac{g'p^{2}}{r^{2}}f_{s}f_{LL}\Phi'' + f_{s}\frac{p^{2}}{r^{2}}\left(b + \frac{pf_{s}}{r}\right)\left(\frac{g''(\Psi')^{2}}{g'} - g'\Psi''\right)\left(\frac{\Phi''}{p} - f_{FF}\right) - \frac{f_{s}g'p^{3}}{r^{2}}\left(f_{FF}f_{ss} - f_{sF}^{2}\right) < 0$$
(18)
$$\frac{dL_{o}}{dr} = \frac{p_{r}f_{ss}g'}{r}f_{ss}g' + \frac{p_{r}}{r^{2}}f_{s}g' + \frac{p_{r}}{r}f_{sF}g' - \frac{p_{r}}{r}g' - \frac{p_{r}}$$

since numerator is negative:

$$-\frac{f_{S}g'p^{3}}{r^{2}}\left[f_{LF}f_{SF}+f_{SL}\left(\frac{\Phi''}{p}-f_{FF}\right)\right] < 0$$

$$(19) \quad \frac{dF}{dr} = \frac{\begin{bmatrix} pf_{FS} & pf_{FL} & 0\\ \left[pf_{LS} - \frac{p}{r}f_{SS}\Psi' \right] & \left[pf_{LL} - \left(b + \frac{p}{r}f_{S}\right)\Psi'' - \frac{p}{r}f_{SL}\Psi' \right] & -\frac{p}{r^{2}}f_{S}\Psi'\\ \frac{p}{r}f_{SS}g' & \left[\left(b + \frac{p}{r}f_{S}\right)\frac{g''}{g'}\Psi' + \frac{p}{r}f_{SL}g' \right] & \frac{p}{r^{2}}f_{S}g' \\ |J| < 0$$

since the sign of the numerator is negative:

$$\frac{f_{s}p^{2}}{r^{3}g'}\left[(br+pf_{s})\underbrace{\left(g''(\Psi')^{2}-(g')^{2}\Psi''\right)}_{(-)}\underbrace{\left(\frac{\Phi''}{p}-f_{FF}\right)}_{(+)}+pr(g')^{2}\underbrace{\left(f_{LL}f_{SF}-f_{LF}f_{SL}\right)}_{(-)}\right]<0$$

$$(20) \quad \frac{dS}{dw} = \frac{\begin{bmatrix} 0 & pf_{FL} & [pf_{FF} - \Phi''] \\ 1 & \left[pf_{LL} - \left(b + \frac{p}{r} f_{S} \right) \Psi'' - \frac{p}{r} f_{SL} \Psi' \right] & \left[pf_{LF} - \frac{p}{r} f_{SF} \Psi' \right] \\ 1 & \left[\left(b + \frac{p}{r} f_{S} \right) \frac{g''}{g'} \Psi' + \frac{p}{r} f_{SL} g' \right] & \frac{p}{r} f_{SF} g' \\ \end{bmatrix}}{|J|} \le 0 \quad or \geq 0$$

since the sign of the numerator is indeterminate:

$$\underbrace{\left(\underbrace{g''}_{g'}\Psi'+\Psi''\right)}_{(?)}\underbrace{\left(\underbrace{f_{FF}}_{(-)}-\Phi''\right)}_{(-)}\left(b+\frac{p}{r}f_{s}\right)-p^{2}\underbrace{\left(\underbrace{f_{FF}}_{LL}-f_{sL}^{2}\right)}_{(+)}+\underbrace{p\Phi''_{f_{LL}}}_{(-)}-\frac{p}{r}\left(g'+\Psi'\right)\underbrace{\left(pf_{LF}f_{SF}-pf_{FF}f_{SL}+\Phi''_{f_{SL}}\right)}_{(+)}\leq 0 \text{ or } \geq 0$$

$$(21) \quad \frac{dL_o}{dw} = \frac{\begin{bmatrix} pf_{FS} & 0 & \left[pf_{FF} - \Phi''\right] \\ \left[pf_{LS} - \frac{p}{r}f_{SS}\Psi'\right] & 1 & \left[pf_{LF} - \frac{p}{r}f_{SF}\Psi'\right] \\ \frac{p}{r}f_{SS}g' & 1 & \frac{p}{r}f_{SF}g' \\ |J| < 0$$

since the sign of the numerator is negative: \Box

$$-\frac{p^{2}}{r}\left[\underbrace{rf_{LF}f_{SF}}_{(+)} + \frac{r}{\underbrace{p}}f_{SL}\Phi'' - \underbrace{rf_{FF}f_{SL}}_{(-)} + \underbrace{\Psi'f_{FF}f_{SS}}_{(+)} + \left(g' + \Psi'\right)\left(\underbrace{\left(f_{FF}f_{SS} - f_{SF}^{2}\right) - \frac{\Phi''f_{SS}}{p}}_{(+)}\right)\right] < 0$$

$$(22) \quad \frac{dF}{dw} = \frac{\begin{bmatrix} pf_{FS} & pf_{FL} & 0\\ \left[pf_{LS} - \frac{p}{r}f_{SS}\Psi' \right] & \left[pf_{LL} - \left(b + \frac{p}{r}f_{S}\right)\Psi'' - \frac{p}{r}f_{SL}\Psi' \right] & 1\\ \frac{p}{r}f_{SS}g' & \left[\left(b + \frac{p}{r}f_{S}\right)\frac{g''}{g'}\Psi' + \frac{p}{r}f_{SL}g' \right] & 1\\ |J| \leq 0 \quad or \geq 0 \end{bmatrix}$$

since the sign of the numerator is indeterminate:

$$\frac{p^{2}}{r}\left[\underbrace{rf_{LL}f_{SF}}_{(-)} - \underbrace{rf_{LF}f_{SL}}_{(+)} - \underbrace{\left(\frac{g''}{g'}\Psi' + \Psi''\right)}_{(?)}\underbrace{\left(bpr \cdot f_{SF} + f_{S}f_{SF}\right)}_{(+)} + \left(g' + \Psi'\right)\underbrace{\left(f_{LF}f_{SS} - f_{SF}f_{SL}\right)}_{(-)}\right] \leq 0 \quad or \geq 0$$

As before,
$$\frac{dF}{dw} < 0$$
 if $\frac{g''}{g'}\Psi' + \Psi'' \ge 0$ or $\left|\frac{g''}{g'}\right| \le \left|\frac{\Psi''}{\Psi'}\right|$.

$$(23) \quad \frac{dS}{dv} = \frac{\begin{bmatrix} 1 & pf_{FL} & [pf_{FF} - \Phi''] \\ 0 & \left[pf_{LL} - \left(b + \frac{p}{r} f_{S} \right) \Psi'' - \frac{p}{r} f_{SL} \Psi' \right] & \left[pf_{LF} - \frac{p}{r} f_{SF} \Psi' \right] \\ \begin{bmatrix} 0 & \left[\left(b + \frac{p}{r} f_{S} \right) \frac{g''}{g'} \Psi' + \frac{p}{r} f_{SL} g' \right] & \frac{p}{r} f_{SF} g' \end{bmatrix}}{|J|} \le 0 \quad or \geq 0$$

since the sign of the numerator is indeterminate: $\[\] \]$

$$\left(b + \frac{pf_s}{r}\right) \left[\underbrace{\frac{g''p\Psi'}{g'}}_{(-)} \underbrace{\left(\frac{\Psi'f_{SF}}{r} - f_{LF}\right)}_{(?)} - \underbrace{\frac{f_{SF}g'p\Psi''}{r}}_{(+)}\right] + \frac{g'p^2}{r} \underbrace{\left(f_{LL}f_{SF} - f_{LF}f_{SL}\right)}_{(-)} \le 0 \quad or \ \ge 0$$

$$\frac{dS}{dv} < 0 \text{ if } \frac{\Psi' f_{SF}}{r} - f_{LF} \ge 0$$

$$(24) \quad \frac{dL_o}{dv} = \frac{\begin{bmatrix} pf_{FS} & 1 & [pf_{FF} - \Phi''] \\ pf_{LS} - \frac{p}{r} f_{SS} \Psi' \end{bmatrix} 0 \begin{bmatrix} pf_{LF} - \frac{p}{r} f_{SF} \Psi' \\ pf_{LS} - \frac{p}{r} f_{SS} g' & 0 & \frac{p}{r} f_{SF} g' \\ \end{bmatrix}}{|J|} < 0$$

since the sign of the numerator is negative:

$$\frac{g'p^2}{\underset{(+)}{r}} \left[\underbrace{f_{SS}f_{LF}}_{(-)} - \underbrace{f_{SF}f_{SL}}_{(+)} \right] < 0$$

$$(25) \quad \frac{dF}{dv} = \frac{\begin{bmatrix} pf_{FS} & pf_{FL} & 1\\ pf_{LS} - \frac{p}{r} f_{SS} \Psi' \end{bmatrix} \begin{bmatrix} pf_{LL} - \left(b + \frac{p}{r} f_{S}\right) \Psi'' - \frac{p}{r} f_{SL} \Psi' \end{bmatrix} 0}{|J|} < 0$$
since the sign of the numerator is negative:

$$\begin{bmatrix} b + \frac{pf_{S}}{r} \end{bmatrix} \begin{bmatrix} \frac{f_{SS}g P \Psi}{r} - \frac{g P \Psi}{g'} \\ \frac{f_{SS}g P \Psi}{r} - \frac{g P \Psi}{g'} \\ \frac{g'}{r} \end{bmatrix} \begin{bmatrix} \frac{f_{SS}g P \Psi}{r} - \frac{g P \Psi}{g'} \\ \frac{g'}{r} \end{bmatrix} \begin{bmatrix} \frac{f_{SS}g P \Psi}{r} - \frac{g P \Psi}{g'} \\ \frac{g'}{r} \end{bmatrix} \begin{bmatrix} \frac{f_{SS}g P \Psi}{r} - \frac{g P \Psi}{r} \\ \frac{g'}{r} \end{bmatrix} \begin{bmatrix} \frac{g' p^{2}}{r} \\ \frac{f_{LL} f_{SS} - f_{SL}^{2}}{r} \end{bmatrix} < 0$$

$$(26) \quad \frac{dS}{dp} = \frac{\left[\left[f_{L} + \frac{f_{S}}{r} \Psi' \right] \\ -\frac{f_{S}}{r} g' \\ \frac{g'}{r} \end{bmatrix} \begin{bmatrix} \left[pf_{LL} - \left(b + \frac{p}{r} f_{S}\right) \Psi'' - \frac{p}{r} f_{SL} \Psi' \right] \\ \frac{f_{SF}g'}{r} \end{bmatrix} \begin{bmatrix} pf_{LL} - \left(b + \frac{p}{r} f_{S}\right) \Psi'' - \frac{p}{r} f_{SL} \Psi' \end{bmatrix} \\ \frac{f_{SF}g'}{r} \end{bmatrix} = 0 \quad or \geq 0$$

since the sign of the numerator is indeterminate: $\[Gamma]$

$$\frac{p^{2}g'}{r}\left[\frac{\Phi''}{p}\underbrace{\left(f_{L}f_{SL}-f_{S}f_{LL}\right)}_{(+)}+f_{S}\underbrace{\left(f_{FF}f_{LL}-f_{FL}^{2}\right)}_{(+)}+f_{L}\underbrace{\left(f_{LF}f_{SF}-f_{FF}f_{SL}\right)}_{(+)}+f_{F}\underbrace{\left(f_{LF}f_{SL}-f_{LL}f_{SF}\right)}_{(+)}\right]+$$

$$+\underbrace{\left(b+\frac{pf_{S}}{r}\right)}_{(+)}\underbrace{\left(\frac{g'}{r}\Psi''-\frac{g''}{rg'}(\Psi')^{2}\right)}_{(+)}\underbrace{\left(pf_{F}f_{SF}-pf_{S}f_{FF}+f_{S}\Phi''\right)}_{(+)}+$$

$$+\underbrace{\frac{g''}{g'}\Psi'\left(b+\frac{pf_{S}}{r}\right)}_{(-)}\underbrace{\left(pf_{F}f_{LF}-pf_{L}f_{FF}+f_{L}\Phi''\right)}_{(-)}\leq 0 \quad or \geq 0$$

$$(27) \quad \frac{dL_{Q}}{dp} = \frac{\begin{bmatrix} pf_{FS} & -f_{F} & [pf_{FF} - \Phi''] \\ pf_{LS} - \frac{p}{r} f_{SS} \Psi' \end{bmatrix} \begin{bmatrix} f_{L} + \frac{f_{S}}{r} \Psi' \end{bmatrix} \begin{bmatrix} pf_{LF} - \frac{p}{r} f_{SF} \Psi' \\ pf_{LS} - \frac{p}{r} f_{SS} \varphi' & -\frac{f_{S}}{r} \varphi' \\ \frac{p}{r} f_{SS} \varphi' & -\frac{f_{S}}{r} \varphi' \\ |J| > 0$$

since the sign of the numerator is positive:

$$\frac{p^2g'}{r}\left[f_s\left(\underbrace{f_{LF}f_{SF} - f_{FF}f_{SL}}_{(+)} + \frac{\Phi''}{p}f_{SL}\right) + f_L\left(\underbrace{f_{FF}f_{SS} - f_{SF}^2}_{(+)} - \frac{\Phi''}{p}\underbrace{f_{SS}}_{(-)}\right) + f_F\underbrace{\left(f_{SF}f_{SL} - f_{LF}f_{SS}\right)}_{(+)}\right] > 0$$

$$\begin{aligned} & \left[\begin{array}{ccc} pf_{FS} & pf_{FL} & -f_{F} \\ \left[pf_{LS} - \frac{p}{r} f_{SS} \Psi' \right] & \left[pf_{LL} - \left(b + \frac{p}{r} f_{S} \right) \Psi'' - \frac{p}{r} f_{SL} \Psi' \right] & \left[f_{L} + \frac{f_{S}}{r} \Psi' \right] \\ & \left[\frac{p}{r} f_{SS} g' & \left[\left(b + \frac{p}{r} f_{S} \right) \frac{g''}{g'} \Psi' + \frac{p}{r} f_{SL} g' \right] & -\frac{f_{S}}{r} g' \\ & \left| J \right| \\ \end{aligned} \right] \\ \end{aligned} \\ (28) \quad \frac{dF}{dp} = \frac{\left[\frac{p}{r} f_{SS} g' & \left[\left(b + \frac{p}{r} f_{S} \right) \frac{g''}{g'} \Psi' + \frac{p}{r} f_{SL} g' \right] & -\frac{f_{S}}{r} g' \\ & \left| J \right| \\ \end{array} \right] \\ \leq 0 \quad or \quad \geq 0 \\ \end{aligned} \\ \frac{pg'}{since} \left(h + \frac{pf_{S}}{sign} \right) \underbrace{\left(\frac{f_{ref} f_{sem} - f_{sf} f_{sf}}{(-)} \right)}_{(-)} + \underbrace{\left(\frac{g''}{pg'} (\Psi')^{2} - \Psi'' \right)}_{(-)} + \underbrace{\frac{pg''\Psi'}{g'}}_{(-)} \left(b + \frac{pf_{S}}{r} \right) \underbrace{\left(f_{L} f_{SF} - f_{F} f_{SL} \right)}_{(7)} + \\ + \frac{pg'}{r} \left[pf_{S} \underbrace{\left(f_{LF} f_{SL} - f_{LL} f_{SF} \right)}_{(+)} + pf_{L} \underbrace{\left(f_{SF} f_{SL} - f_{LF} f_{SS} \right)}_{(+)} + pf_{F} \underbrace{\left(f_{LL} f_{SS} - f_{SL}^{2} \right)}_{(+)} \right] \\ \leq 0 \quad or \quad \geq 0 \\ \underbrace{\frac{dF}{dp}} > 0 \quad \text{if} \quad f_{L} f_{SF} - f_{F} f_{SL} \leq 0 \quad \text{or} \quad \underbrace{\frac{\partial \left(\frac{f_{F}}{f_{L}} \right)}_{\partial S} \leq 0}_{\partial S} \\ \leq 0, \text{ i.e. if an increase in soil capital increases} \end{aligned}$$

marginal product of labour more than marginal product of fertilizer.

References

Admassie, Yeraswerq, 1992. *The Catchment Approach to Soil Conservation in Kenya*, Technical Report nr 6, Sida's Regional Land Management Unit, Nairobi.

Alemu, Tekie, 1999. *Land Tenure and Soil Conservation: Evidence from Ethiopia*, (Ph. D. thesis), Department of Economics, Göteborg University.

Anderson, Donald M., 1995. *Toxic Red Tides and Harmful Algal Blooms: A Practical Challenge in Coastal Oceanography*, **Reviews of Geophysics**, Vol. 33, Iss. S1, p. 1189-1200.

Andersson, Jessica, 2004. *Welfare, Environment and Tourism in Developing Countries* (Ph. D. thesis), Department of Economics, Göteborg University.

Ayoub, Ali T., 1999. *Fertilizers and the Environment*, **Nutrient Cycling in Agroecosystems**, Vol. 55, Nr. 2, p. 117-121, Springer Science.

Ballot, Andreas, Lothar Krienitz, Kiplagat Kotut, Claudia Wiegand, James S. Metcalf, Geoffrey A. Codd and Stephan Pflugmacher, 2004. *Cyanobacteria and Cyanobacterial Toxins in Three Alkaline Rift Valley Lakes of Kenya - Lakes Bogoria, Nakuru and Elmenteita*, **Journal of Plankton Research**, Vol. 26,Iss. 8, p. 925-935, Oxford University Press

Barbier, Edward, 1990. *The Farm-Level Economics of Soil Conservation: The Uplands of Java*, Land Economics 66(2):199-211.

Bartram, Jamie and Ingrid Chorus (Ed.), 1999. *Toxic Cyanobacteria in Water: A guide to their Public Health Consequences, Monitoring and Management,* World Health Organization. http://www.who.int/docstore/water_sanitation_health/toxicyanobact/begin.htm#Contents

Bartram Jamie, Kristen Lewis, Roberto Lenton and Albert Wright, 2005. *Millennium Project : Focusing on Improved Water and Sanitation for Health*, **The Lancet**, Vol. 365, p. 810-12, Elsewier.

Bastidas, C., D. Bone and E. M. García, 1999. Sedimentation Rates and Metal content of Sediments in a Venezuelan Coral Reef, Marine Pollution Bulletin, Vol. 38, Iss. 1, p. 16-24.

Barrett, Scott, 1991. Optimal Soil Conservation and the Reform of Agricultural Pricing Policies, Journal of Development Economics, Vol. 36, Iss. 2, p. 167-187, Elsevier Science

Evenson, Robert E. and G. Mwabu, 2001. *The Effect of Agricultural Extension on Farm Yields in Kenya*, African Development Review, Vol. 13, Nr. 1, p. 1-23

Bishop, J. and Allen, J. 1989. *The On-Site Costs of Soil Erosion in Mali*, Environment Dept. Working Paper No. 21, The World Bank, Washington DC.

Bryant, D.G., L. Burke, J. McManus and M. Spalding, 1998. *Reefs at Risk: a Map-based Indicator of Threats to the World's Coral Reefs*, World Resources Institute, Washington, DC.

Bunce, A., 1942. The Economics of Soil Conservation, Ames: Iowa State College Press.

Burt, Oscar R., 1986. Farm Level Economics of Soil Conservation in the Palouse Area of the Northwest: Reply, American Journal of Agricultural Economics, Vol. 68, Iss. 2, p. 367-69

Burt, Oscar R., 1981. *Farm Level Economics of Soil Conservation in the Palouse Area of the Northwest*, **American Journal of Agricultural Economics**, Vol. 63, Iss. 1, p. 83-92

Campos, José, J., Francisco Alpizar, Bastiaan Louman and John Parotta, 2005. An Integrated Approach to Forest Ecosystem Services in Forests in the Global Balance - Changing Paradigms (ed.) Campos et al., UIFRO.

Christiansson, Carl, 1991. Use and Impacts of Chemical Pesticides in Smallholder Agriculture in the Central Kenya Highlands, In Linking the Natural Environment and the Economy (ed.) Carl Folke and Tomas Kåberger, Kluwer Academic Publishers, Dordrecht.

Clark, E.H., J.A. Haverkamp, and W. Chapman, 1985. *Eroding Soils: The Off-Farm Impacts*, The Conservation Foundation, Washington DC.

Clarke, Harry R., 1992. *The Supply of Non-degraded Agricultural Land*, Australian Journal of Agricultural Economics, v. 36, iss. 1, pp. 31-56.

Cortes J.N. and M.J. Risk, 1985. A Reef Under Siltation Stress: Cahuita, Costa Rica, Bulletin of Marine Science, Vol. 36, p. 339–356.

Costa Jr., O.S., Z.M. Leao, M. Nimmo and M.J. Attrill, 2000. *Nutrification impacts on Coral Reefs from Northern Bahia, Brazil*, Hydrobiologia, Vol. 440, p. 370–415, Elsevier Ltd.

Crosson, Pierre, 1986. *Soil Erosion and Policy Issues*, in Phillips, T., P. Crosson, and K. Price (eds) <u>Agriculture and the Environment</u>. Resources for the Future, Washington DC.

Cruz, W., Francisco, H.A. and Conway, Z.T. 1988. *The On-Site and Downstream Costs of Soil Erosion in the Magat and Pantabangan Watersheds*, J. of Philippine Development 15(26, 1), p. 85-111.

Cuesta, M., Carlson, G and Ernst Lutz, 1997. *An Empirical Estimation of Farmers' Discount Rates in Costa Rica*, The World Bank, Washington DC.

Dorfman, Robert, 1969. An Economic Interpretation of Optimal Control Theory, The American Economic Review, Vol. 59, No. 5. (Dec., 1969), pp. 817-831.

Ekbom, Anders, 2007. *Determinants of Soil Capital - an Application to Kenya's Central Highlands*, (mimeo) Department of Economics with Statistics, Göteborg University

Ekbom, Anders, Per Knutsson and Mira Ovuka, 2001. *Is Sustainable Development Based on Agriculture Attainable in Kenya?* J. of Land Degradation and Development, Vol. 12, pp. 435-447, John Wiley & Sons.

El-Swaify, S. A., 1997. Factors Affecting Soil Erosion Hazards and Conservation Needs for Tropical Steeplands, Soil Technology, Vol. 11, Iss. 1, pp. 3-16, Elsevier B.V.

Fabricius, Katharina E., and G. De'ath, 2001. *Environmental Factors Associated with the Spatial Distribution of Crustose Coralline Algae on the Great Barrier Reef*, Coral Reefs Vol. 19, p. 303–309.

Fabricius, Katharina E., 2004. *Effects of Terrestrial Runoff on the Ecology of Corals and Coral Reefs: Review and Synthesis*, **Marine Pollution Bulletin**, (in press, Corrected Proof, Available online 9 December 2004, Elsevier Science Ltd.

FAO (UN Food and Agriculture Organization), 1995. *World Agriculture: Towards 2010*, [Ed. N. Alexandratos], FAO, Rome.

Feder, Gershon and David Feeny, 1991. Land Tenure and Property Rights: Theory and Implications for Development Policy, World Bank Economic Review, Vol. 5, Iss. 1, p. 135-53.

Feder, Gershon and Tongroj Onchan, 1987. Land Ownership Security and Farm Investment in Thailand, American Journal of Agricultural Economics, Vol. 69, Iss. 2, p. 311-20.

Fernex, François, Pedro Zárate-del Valle, Hermes Ramírez-Sánchez, François Michaud, Claude Parron, Josette Dalmasso, Geneviève Barci-Funel and Manuel Guzman-Arroyo, 2001. *Sedimentation rates in Lake Chapala (western Mexico): possible Active Tectonic Control*, **Chem. Geology**, Vol. 177, Iss. 3-4, p. 213-228.

Gachene, Charles, 1995. Effect of Soil Erosion on Soil Properties and Crop Response in Central Kenya (PhD thesis), Department of Soil Sciences, Swedish University of Agricultural Sciences, Uppsala.

Gachene, Charles, Nicholas Jarvis, H. Linner, J.P. Mbuvi, 1997. Soil Erosion Effect on Soil Properties in a Highland of Central Kenya, Soil Sciences Society of America Journal, Vol. 61, pp. 559-564.

Gautam, Madhur and Jock R. Anderson, 1999. *Reconsidering the Evidence on the Returns to T&V Extension in Kenya*, Policy Research Working Paper No 2098, Operations Evaluations Department, World Bank.

Goetz, Renan U., 1997. *Diversification in Agricultural Production: A Dynamic Model of Optimal Cropping to Manage Soil Erosion*; American Journal of Agricultural Economics, v. 79, iss. 2, pp. 341-56, American Agricultural Economics Association.

Golbuu, Yimnang, Steven Victor, Eric Wolanski and Robert Richmond, 2003. *Trapping of Fine Sediment in a Semi-Enclosed Bay, Palau, Micronesia*, **Estuarine, Coastal and Shelf Science**, Vol. 57, Iss. 5-6, p. 941-949

Grepperud, Sverre, 2000. *Optimal Soil Depletion with Output and Price Uncertainty*, J. of Environment and Development Economics, v. 5, iss. 3, pp. 221-40, Cambridge, Cambridge Univ. Press

Grepperud, Sverre, 1997a. *Soil Conservation as an Investment in Land*, Journal of Development Economics, Vol. 54, Iss. 2, p. 455-67.

Grepperud, Sverre, 1997b. *Poverty, Land Degradation and Climatic Uncertainty*, **Oxford Economic Papers**, Vol. 49, Iss. 4, p. 586-608

Grepperud, Sverre, 1996. *Population Pressure and Land Degradation: The Case of Ethiopia*, Journal of Environmental Economics and Management, Vol. 30, Iss. 1, p. 18-33

Gutman, Pablo, 2003. From Goodwill to Payments for Environmental Services: A Survey of Financing Options for Sustainable Natural Resource Management in Developing Countries, WWF Macroeconomics Program Office, Washington D.C.

Hamed, Yasser, Jean Albergel, Yannick Pépin, Jean Asseline, Slah Nasri, Patrick Zante, Ronny Berndtsson, Mohamed El-Niazy and Mohamed Balah, 2002. *Comparison between Rainfall Simulator Erosion and Observed Reservoir Sedimentation in an Erosion-Sensitive Semiarid Catchment*, CATENA, Vol. 50, Iss. 1, p. 1-16, Elsevier.

Healy, T., Y. Wang and J. A. Healy (eds.), 2002. *Muddy Coasts of the World: Processes, Deposits and Functions*, Elsevier, Amsterdam.

Hodgson, Gregor, 1990. Sediment and the Settlement of Larvae of the Reef Coral Pocillopora damicornis, Coral Reefs, Vol. 9, p. 41–43.

Holden, Stein T. and Nadarajah Shanmugaratnam, 1995. *Strucutral Adjustment, Production Subsidies and Sustainable Land Use*, Forum for Development Studies, Nr. 2, p. 247-266.

Holmes, K.E., E.N. Edinger, Hariyadi, G.V. Limmon and M.J., 2000. *Risk, Bioerosion of Live Massive Corals and Branching Coral Rubble on Indonesian Coral Reefs*, **Marine Pollution Bulletin** Vol. 7, p. 606–617.

Holmes, Thomas P.,1988. *The Offsite Impact of Soil Erosion on the Water Treatment Industry*, Land Economics, November 1988, v. 64, iss. 4, pp. 356-66

Horner, RA, D.L. Garrison and F.G. Plumley, 1997. *Harmful Algal Blooms and Red Tide Problems on the U.S. West Coast*, Limnology and Oceanography, Vol. 42, no. 5, p. 1076-1088.

Hubbard, D.K., 1986. Sedimentation as a Control of Reef Development: St. Croix, U.S.V.I, Coral Reefs, Vol. 5, p. 117–125.

Hudson, Norman, 1985. Soil Conservation, Batsford Academic and Educ. Publishers, Batsford, UK.

De Janvry, Alain, Elisabeth Sadoulet and Blas Santos, 1995. *Project Evaluation for Sustainable Development: Plan Sierra in the Dominican Republic*, Journal of Environmental Economics and Management, Vol 28, Iss. 2, p. 135-154.

van Katwijk, M., N. Meier, R. van Loon, E. van Hove, W Giesen, G. van der Velde and D. den Hartog, 1993. Sabaki River Sediment Load and Coral Stress: Correlation between Sediments and Condition of the Malindi Watamu reefs in Kenya (Indian Ocean). Marine Biology, Vol. 117, p. 675–683.

Kerr, J., 2002. *Watershed Development, Environmental Services, and Poverty Alleviation in India*, **World Development,** Vol. 30, Iss. 8, p. 1387-1400.

Kiker, C. and Lynne, G. 1986. An Economic Model of Soil Conservation: Comment, American Journal of Agricultural Economics 68(3):739-742.

Kimaru, Gathiru, 1998. Kenya's National Soil and Water Conservation Programme (NSWCP), Danish Institute for International Studies, Department for Development Research.

Kosoy, Nicolas and Miguel Martinez-Tuna, Roldan Muradian and Joan Martinez-Alier, 2006. *Payments for Environmental Services in Watersheds: Insights from a Comparative study of Three Cases in Central America*, **Ecological Economics**, Vol. 61, Iss. 2-3, p. 446-455.

LaFrance, Jeffrey T., 1992. Do Increased Commodity Prices Lead to More or Less Soil Degradation?, Australian Journal of Agricultural Economics, v. 36, iss. 1, pp. 57-82.

Lal, Rattan, Pierce, F., and Dowdy R., 1983. *The Threat of Soil Erosion to Long-term Crop Production*, Science, 219, pp 458-465.

Lal, Rattan. 1987. *Effects of erosion on crop productivity*, Critical Reviews in Plant Sciences 5(4):303-367.

Lal, Rattan, 1995. Erosion-Crop Productivity Relationships of Africa, Soil Science of America Journal, Vol. 59, p. 667-667.

Landell-Mills, N., and I. Porras. 2002. Silver Bullet or Fools' Gold? A Global Review of Markets for Forest Environmental Services and Their Impact on the Poor, IIED, London.

Lewis, Lawrence A., 1985. Assessing Soil Loss in Kiambu and Murang'a Districts, Kenya, Geogr. Ann., Vol 67A, pp 273-284.

Loya, Y., 1976. Effects of Water Turbidity and Sedimentation on the Community Structure of Puerto Rican Corals, Bulletin of Marine Science Vol. 26, p. 450–466, Elsevier Ltd.

Lundgren, L., 1993. *Twenty Years of Soil Conservation in Eastern Africa*. Regional Soil Conservation Unit, Report No. 9, Majestic Printing Works, Nairobi, Kenya.

Magrath, William and Peter Arens, 1989. *The Costs of Soil Erosion on Java: A Natural Resource Accounting Approach*, Environment Department Working Paper nr. 18, The World Bank.

Mahmood, K., 1987. Reservoir Sedimentation: Impact, Extent and Mitigation, World Bank Technical Paper, Washington DC.

Matson, P. A., W. J. Parton, A. G. Power, M. J. Swift, 1997. *Agricultural Intensification and Ecosystem Properties*, Science, Vol 277, Iss. 5325, p- 504-509.

McClanahan, T. R. and D. Obura, 1997. Sedimentation Effects on Shallow Coral Communities in Kenya, Journal of Experimental Marine Biology and Ecology, Vol. 209, Iss. 1-2, p. 103-122.

McConnell, Kenneth E. 1983. *An Economic Model of Soil Conservation*, American Journal of Agricultural Economics, Vol. 65, Nr. 1, p. 83-89, American Agricultural Economics Association.

McConnell, Kenneth E. 1986. *An Economic Model of Soil Conservation: A Reply,* American Journal of Agricultural Economics 68(3):743-744, American Agricultural Economics Association.

Miranowski, John A., 1984. Impacts of Productivity Loss on Crop Production and Management in a Dynamic Economic Model, American Journal of Agricultural Economics, v. 66, iss. 1, pp. 61-71, American Agricultural Economics Association

Moore, Walter B., and Bruce A. MacCarl, 1987. *Off-Site Costs of Soil Erosion: A Case Study in the Willamette Valley*, Western Journal of Agricultural Economics, Vol 12 (July), pp. 42-49.

Morgan, Roy P. C., 1986. Soil Erosion and Conservation, John Wiley & Sons, Inc., New York.

Morris, G.L. and J. Fan, 1998. Reservoir Sedimentation Handbook: Design and Management of Dams, reservoirs and Watersheds for Sustainable Use, McGraw Hill, New York.

Nagle, Gregory N., 2002. The Contribution of Agricultural Erosion to Reservoir Sedimentation in the Dominican Republic, Water Policy, Vol. 3, Iss. 6, p. 491-505, Elsevier Science Ltd.

Nearing, M. A., M. J. M. Romkens, L. D. Norton, D. E. Stott, F. E. Rhoton, J. M. Laflen, D. C. Flanagan, C. V. Alonso, R. L. Binger, S. M. Dabney, O. C. Doering, C. H. Huang, K. C. McGregor, and A. Simon; 2000. *Measurements and Models of Soil Loss Rates*, Science November 17; 290: pp.1300-1301.

Naidu R., S. Baskaran, R.S. Kookana, 1998. *Pesticide Fate and Behaviour in Australian Soils in Relation to Contamination and Management of Soil and Water: A Review*, Australian Journal of Soil Research, Vol. 36, Nr. 5, p. 715-764, CSIRO Publishing.

Ochieng, William R. and Robert. M. Maxon [Ed.], 1992. An Economic History of Kenya, East African Educational Publishers Ltd., Nairobi.

OPTO, 2006: Impact Assessment of the National Agriculture and Livestock Extension Programme (NALEP) Phase I, OPTO International AB.

Otieno, F.A.O. and S.M. Maingi. 1993. Sedimentation Problems of Masinga Reservoir, paper presented at the 4th Land and Water Management Workshop Feb. 15-19, 1993, Nairobi, Kenya

Orazem, Peter F. and John A. Miranowski, 1994. *A Dynamic Model of Acreage Allocation with General and Crop-Specific Soil Capital*, **American Journal of Agricultural Economics**, v. 76, iss. 3, pp. 385-95, American Agricultural Economics Association

Ovuka, Mira and Anders Ekbom¹ 1999; *Farmers' Resource Levels, Soil Properties and Productivity in Kenya's Central Highlands*, in *Sustaining the Global Farm* by D.E. Stott, R.H. Mohtar and G.C. Steinhardt (Eds.) 2001, pp. 682-687, Selected scientific papers from the 10th International Soil Conservation Organization meeting May 24-29, 1999 at Purdue University and USDA-ARS National Soil Erosion Research Laboratory.

Pagiola, Stefano and Gunars Platais, 2002. *Payments for Environmental Services*, World Bank/IBRD, Washington D.C.

Pagiola, S., A. Arcenas, and G. Platais. 2005. *Can Payments for Environmental Services help Reduce Poverty? An Exploration of the Issues and the Evidence to Date from Latin America*, World **Development**, Vol. 33, p. 237–253.

Pagiola, S., N. Landell-Mills, and J. Bishop. 2002. *Making Market-based Mechanisms work for Forests and People*, <u>in</u> S. Pagiola, J. Bishop, and N. Landell-Mills (eds.), *Selling Forest Environmental Services: Market-based Mechanisms for Conservation and Development*. Earthscan, London.

Pagiola, Stefano, 2007. *Payments for Environmental Services in Costa Rica*, Ecological Economics (accepted July 27, 2007; forthcoming)

Palmieri, Alesssandro, Farhed Shah F., George W. Annandale and Ariel. Dinar, 2003. *Reservoir Conservation: Economic and Engineering Evaluation of Alternative Strategies for Managing Sedimentation in Storage Reservoirs; The RESCON Approach Vol. I. World Bank, Washington DC.*

Palmieri, Alesssandro, Farhed Shah F. and Ariel Dinar, 2001. *Economics of Reservoir Sedimentation and Sustainable Management of Dams*, Journal of Environmental Management, vol. 61, no. 2, p. 149-163, Academic Press.

Pimentel, David and Edward L. Skidmore, 1999. Rates of Soil Erosion, Science Vol. 19; Iss 286, p. 1477

Polous, C and Dale Whittington, 1999. *Individuals' Time Preference for Life-Saving Programs: Results from Six Less Developed Countries* (mimeo), Dept of Environmental Science and Engineering, University of North Carolina at Chapel Hill.

Pontryagin, L., V. Boltyanskii, R. Gamkrelidze and E. Mishchenko, 1964. *The Mathematical Theory of Optimal Processes*, (translation by D.E. Brown), Pergamon press

Renard, K.G., G.R. Foster, G.A. Weesies, D.K. McCool and D.C. Yoder, 1997. *Predicting Soil Erosion by Water: A Guide to Conservation Planning with the Revised Universal Soil Loss Equation (RUSLE)*, <u>Agricultural Handbook 703</u>, Agricultural Research Service, US Department of Agriculture, Washington DC.

Ribaudo, Marc O., 1986. Consideration of Off-Site Impacts in Targeting Soil Conservation Programs, Land Economics, Vol. 62 (Nov.), p. 402-411.

Ross, Jamie and Robert Gilbert, 1999. Lacustrine Sedimentation in a Monsoon Environment: the Record from Phewa Tal, Middle Mountain Region of Nepal, Geomorphology, Vol. 27, Iss. 3-4, p. 307-323

Rowan, J. S., L. E. Price, C. P. Fawcett and P. C. Young, 2001. *Reconstructing Historic Reservoir Sedimentation Rates Using Data-Based Mechanistic Modelling*, **Phys. Chem. Earth** (B), Vol. 26, Iss. 1, p. 77-82, Pergamon.

Saenyi, Wycliffe Wanyonyi, and M. C. Chemelil, 2003. *Modelling of Suspended Sediment Discharge for Masinga Catchment Reservoir in Kenya*, Journal of Civil Engineering, Vol. 8, p. 89-98.

Sanders, David W., Paul C. Huszar, Samran Sombatpanit and Thomas Enters, 1999. *Incentives in Soil Conservation: From Theory to Practice*, World Association of Soil and Water Conservation.

Schneider, H.M., 1993. *Sediment Sources to the Masinga Dam*, paper presented at the 4th Land and Water Management Workshop Feb. 15-19, 1993, Nairobi, Kenya

Shiferaw; Bekele and Stein T. Holden, 2000. Policy Instruments for Sustainable Land Management: the Case of Highland Smallholders in Ethiopia, Agricultural Economics, Vol 22, p. 217-232, Elsevier.

Shiferaw; Bekele and Stein T. Holden, 2001. Farm-level Benefits to Investments for Mitigating Land Degradation: Empirical Evidence from Ethiopia, J. of Environment and Development Economics Vol 6, p. 335-358, Cambridge University Press.

Shimoda, T., T. Ichikawa and Y. Matsukawa, 1998. *Nutrient Conditions and their Effects on Coral Growth in Reefs around Ryukyu Islands*, **Bull. National Research Institute of Fisheries Science** Vol. 12, p. 71–80. Shone, Ronald, 1999. *Economic Dynamics*, Cambridge University Press, Cambridge, UK.

Shumway, S. E., 1990. A Review of the Effects of Algal Blooms on Shellfish and Aquaculture, J. World Aquacult. Soc., Vol. 21, p. 65-104.

Sierra, Rodrigo and Eric Russman, 2005. *On the Efficiency of Environmental Service Payments: A Forest Conservation Assessment in the Osa Peninsula, Costa Rica*, Ecological Economics, Vol. 59, Iss. 1, p. 131-141.

Smith, Elwin G., Mel Lerohl, Teklay Messele and Henry H. Janzen, 2000. *Soil Quality Attribute Time Paths: Optimal Levels and Values*, **Journal of Agricultural and Resource Economics**, Vol. 25, Iss. 1, p. 307-24, American Agricultural Economics Association.

Smith, Elwin G.; Shaykewich, Carl F., 1990. *The Economics of Soil Erosion and Conservation on Six Soil Groupings in Manitoba*, Canadian Journal of Agricultural Economics, Vol. 38, Iss. 2, p. 215-31.

Smith, V. Kerry, 1992. *Environmental Costing for Agriculture: Will It Be Standard Fare in the Farm Bill of 2000?* American Journal of Agricultural Economics, Vol. 74, Iss. 5, p. 1076-88.

Smith, Katherine R., 2006. *Public Payments for Environmental Services from Agriculture: Precedents and Possibilities*, Amer. J. Agr. Econ., Vol. 88, Nr. 5, p. 1167–1173.

Southgate, Douglas and Macke, R., 1989. *The Downstream Benefits of Soil Conservation in Third World Hydroelectric Watersheds*, Land Economics Vol. 65, Iss. 1, p. 38-48.

Sterner, Thomas, 2003. *Policy Instruments for Environmental and Natural Resource Management*, Resources For the Future, RFF Press, Washington D.C.

Stimson J. and S.T. Larned, 2000. Nitrogen Efflux from Sediments of a Subtropical bay and the Potential Contribution to Macro-algal Nutrient Requirements, Journal of Experimental Marine Biology and Ecology, Vol. 252, p. 159–180. Knut Sydsaeter and Peter J. Hammond, 1995. Mathematics for Economic Analysis, Prentice-Hall Inc., New Jersey.

Thomas, Michael F., 1994. *Geomorphology in the Tropics: a study of Weathering and Denudation in Low Latitudes*, John Wiley and Sons, Ltd., England.

Tomascik, T. and F. Sander, 1987. *Effects of Eutrophication on Reef-building Corals: Structure of Sclerac-tinian Coral Communities on Fringing Reefs, Barbados, West Indies, Marine Biology*, Vol. 94, p. 53–75.

Trimble, Stanley W. and Pierre Crosson, 2000. *Measurements and Models of Soil Loss Rates*, Science November 17; 290: pp.1300-1301.

Troeh, Frederick, J. Arthur Hobbs and Roy L. Donahue, 1991. *Soil and Water Conservation for Productivity and Environmental Protection* (2nd ed.), Prentice Hall, Englewood Cliffs NJ.

USACE, 1985. *Beneficial Uses of Dredged Material*, US Department of the Army Corps of Engineers (USACE), EM1110-2-5026.

de Vente, Joris, Jean Poesen and Gert Verstraeten, 2004. The Application of Semi-Quantitative Methods and Reservoir Sedimentation Rates for the Prediction of Basin Sediment Yield in Spain, **Journal of Hydrology**, October 2004, p. 1-24, Elsevier Science B.V.

White, Alan T., Helge P. Vogt and Tijen Arin, 2000. *Philippine Coral Reefs under Threat: The Economic Losses Caused by Reef Destruction*, Marine Pollution Bulletin, Vol. 40, Iss. 7, p. 598-605, Elsevier Ltd.

White, P., D. P. Butcher and J. C. Labadz, 1997. *Reservoir Sedimentation and Catchment Sediment Yield in the Strines catchment, U.K.*, Phys. Chem. Earth, Vol. 22, Iss. 3-4, p. 321-328, Pergamon

WHO, 2005. Water Sanitation and Health http://www.who.int/water sanitation health/diseases/en/

Wilcox, Walter W., 1938. *Economic Aspects of Soil Conservation*, Journal of Political Economy, Vol. 46, Iss. 5 (Oct 1938), pp. 702-713.

Wischmeier W. H. and Smith D.D., 1965. *Predicting Rainfall Erosion Losses from Cropland East of the Rocky Mountains*, <u>Agricultural Handbook no. 282</u>, U.S. Department of Agriculture, Washington D.C.

Wischmeier W. H. and Smith D.D., 1978. *Predicting Rainfall Erosion Losses*, <u>Agricultural Handbook</u> no. 536, United States Department of Agriculture, Washington D.C.

Wunder, Sven, 2005. *Payments for Environmental Services: Some Nuts and Bolts*, Center for International Forestry Research (CIFOR), Jakarta

Yesuf, Mahmud, 2004. *Risk, Time and Land Management under Market Imperfections: Applications to Ethiopia* (Ph.D. thesis), Department of Economics, Göteborg University.

Younes M. and Jamie Bartram, 2001. *Waterborne Health Risks and the WHO Perspective*, **International Journal of Hygiene and Environmental Health**, Vol. 204, Iss. 4, p. 255-263.

Zbinden, Simon and David R. Lee, 2005. *Paying for Environmental Services: An Analysis of Participation in Costa Rica's PSA Program*, **World Development**, Vol. 33, Iss. 2, p. 255-272.

List of variables (supplementary info for referees)

- Q = Crop output
- F = Fertilizer input
- σ = Net soil loss
- S = Soil capital
- E = Downstream environmental quality
- $b(\dot{s}) = External flow benefit (or cost) of soil motion$
- v = Price of fertilizer
- r = Interest rate
- L_Q = Labour supply to agricultural production
- L_C = Labour supply to soil conservation
- w = Labour wage rate
- f() = Agricultural production fcn.
- p = Crop price
- π = Agricultural profit
- λ = Shadow value of soil
- s = Subsidy to soil conservation
- τ = Pollution charge